

# Bisphenol-A analogue (bisphenol-S) exposure alters female reproductive tract and apoptosis/oxidative gene expression in blastocyst-derived cells

Alireza Nourian<sup>1</sup>, Ali Soleimanzadeh<sup>2\*</sup>, Ali Shalizar Jalali<sup>3</sup>, Gholamreza Najafi<sup>3</sup>

<sup>1</sup> Faculty of Veterinary Medicine, Urmia University, Urmia, Iran

<sup>2</sup> Department of Theriogenology, Faculty of Veterinary Medicine, Urmia University, Urmia, Iran

<sup>3</sup> Department of Basic Sciences, Faculty of Veterinary Medicine, Urmia University, Urmia, Iran

## ARTICLE INFO

**Article type:**  
Original article

**Article history:**  
Received: Jun 3, 2019  
Accepted: Nov 9, 2019

**Keywords:**  
Apoptosis  
Bisphenol-S  
Female reproductive tract  
*In vitro* fertilization  
Oxidative stress

## ABSTRACT

**Objective(s):** One of the major endocrine-disrupting chemicals, bisphenol-S (BPS) has replaced bisphenol-A due to public health anxiety. The present study evaluated low dosage BPS effect on female reproductive potential, hormonal disruption, and gene expression pathways of blastocyst-derived cells.

**Materials and Methods:** NMRI female mice (5-6 weeks) in the estrous stage were chosen following vaginal smear examination for estrus cycle detection and BPS (0, 1, 5, 10, 50 and 100 µg/kg) was administrated subcutaneously for twenty-one consecutive days. After the last administration, blood, ovary tissue and oocytes were collected for further examination.

**Results:** BPS induced oxidative stress in ovarian tissue and reduced hormonal status, LH and FSH, even at low concentration. Furthermore, apoptosis was induced in blastocyst derived cells in BPS administrated mice groups even at low BPS concentration, however, P53 and E2f1 expression were downregulated in doses more than 50 µg/kg, which might indicate apoptosis pathway exchange from P53 dependent to p53 independent pathways. IVF outcome was negatively associated with blastocyst apoptosis gene expression, estrogen receptor beta (ERβ) as well as oxidative status in ovaries. Finally, Stepwise regression indicated that E2f1, Nrf2, catalase (CAT), and malondialdehyde (MDA) could be chosen as predictor values for hatch percentage in IVF outcome.

**Conclusion:** In summary, this study revealed BPS might have detrimental potential in the female reproductive system by oxidation induction and hormonal alteration as well as next generation blastocyst derived cells apoptosis induction. Further studies are recommended for public health assurance of BPS safety especially for female consumed products.

► Please cite this article as:

Nourian AR, Soleimanzadeh A, Shalizar Jalali A, Najafi GhR. Bisphenol- A analogue (bisphenol-S) exposure alters female reproductive tract and apoptosis/oxidative gene expression in blastocyst-derived cells. Iran J Basic Med Sci 2020; 23:576-585. doi: 10.22038/IJBMS.2020.40893.9664

## Introduction

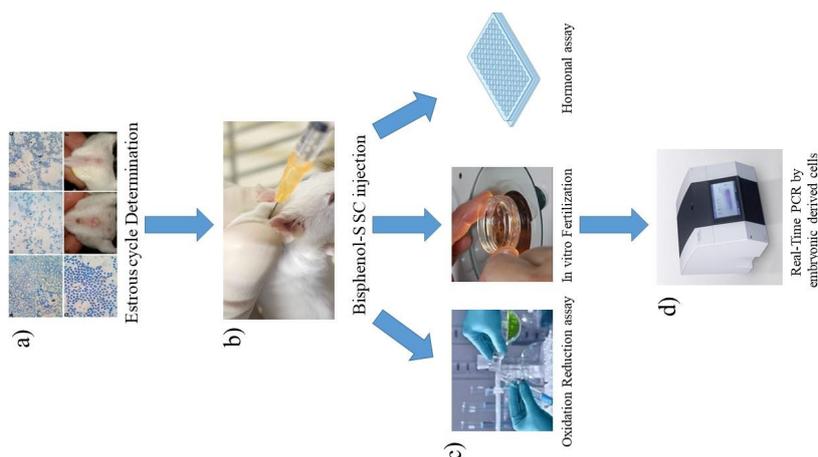
Modern lifestyle has increased human exposure to a variety of environmental toxins that enter the body via respiration and digestion (1). Endocrine disrupting chemicals (EDCs) are the major environmental pollutants that have been used routinely as the main component of many products such as plastics, food containers, cans, and pesticides (2).

Bisphenol-A (BPA) is one of the main EDCs that is widely used as an ingredient of hard polycarbonate plastics, epoxy resins and many other materials. BPA mimics an estrogenic effect by binding competitively to different types of estrogen receptors (3). Many studies investigated that not only BPA had detrimental potential for male and female reproductive system, but also, it decreased thyroid hormone level in serum, reduced hormone production (4, 5), reduced hypothalamic steroid hormone release (6), decreased male gonadotropin hormone level (7), abnormal embryonic development, and implantation induction (8, 9). Similarly, long-term administration of BPA disrupted the estrous cycle of non-pregnant mice (10) and ovarian reserve (11). Furthermore, another study revealed that

low doses of BPA *in vitro* induced oxidative stress in testis (12-14). Due to public health concerns related to the toxic effects of BPA, its application is restricted especially in the US and replaced by "BPA-Free" products that contain substitutes such as bisphenol-F, bisphenol-B and bisphenol-S (BPS) (15).

As the most well-known BPA substitute, BPS is widely used in BPA-Free labeled products. After industrial replacement of BPA by BPS, the annual consumption of BPS has increased to more than 10,000 tons in Europe (16). Average absorption of BPS through foods was evaluated in a population living near the river and it was revealed that the estimated daily intake of BPS was 9 ng/kg of body weight (17). Recent investigations have detected the increased usage of BPS content materials in various countries which led to new concerns about health risks of BPS (18, 19). Recent studies show that BPS is more resistant in the ecosystem than mammalian organelles which increases its hazardous potential compared to BPA (20).

BPS exposure hazardous influence on the reproductive system was detected recently. BPS influenced uterine weight increase, prenatal ovarian follicle development



**Figure 1.** Schematic picture of the experiment. a) After estrus cycle detection all animals were divided randomly and b) BPS administrated Sc. c) IVF, Oxidative assay and ELISA were performed and d) Blastocyst cells harvested for real-time PCR

(21), maternal behavior, and lactation period (22, 23). Moreover, reduction in gonad weight, egg production, hatchability and embryo's stages of development were reported in zebra fish that were exposed to low BPS and BPA dosages (24, 25). Alteration in hormonal activity and hypothalamic-pituitary-gonadal axis in males and females, even at the early stages of development were reported in exposure to different BPS concentrations (25-27). Furthermore, BPS obesity induction was reported in first and second generations by influencing the hypothalamic appetite pathway and deregulating lipid metabolism (20, 28, 29). Different metabolic pathways were detected in infants and pregnant mothers exposed to BPS that caused BPS metabolites abandonment in the infant body and could not be eliminated from them (30). In fact, BPS metabolized by conjugative rather than oxidative metabolism pathways (31). Moreover, an *in vitro* study revealed that oocyte meiosis maturation was interrupted by BPS even at doses lower than the dose human beings are exposed to the environment (32), and the effect of BPS on different embryonic stages was also reported in zebrafish (24).

Like BPA, BPS induced oxidation-reduction at different doses in many tissues such as liver, kidneys, and testes. Moreover, influence of BPS on the male reproductive system was determined in different laboratory conditions (*in vivo* and *in vitro*), where BPS induced detrimental effects by increasing oxidative stress and decreasing testosterone concentration and daily sperm production (13, 33). Furthermore, prenatal exposure to BPS induced steroidogenic enzyme activity in female infants as well as oxidative stress and apoptosis in the male testis of next generation (34).

BPA also induced reproductive system apoptosis by altering different cellular pathways (35, 36). To the best of our knowledge, no study was conducted to determine the detrimental effects of BPS on DNA and apoptosis in the female reproductive system. Based on our previous studies, BPS has a detrimental effect on *in vivo* fertilization capability as well as *in vitro* which was demonstrated by increasing type I arrest of IVF outcome (37, 38).

This study was conducted to illustrate the effect of low concentration BPS on ovarian tissue oxidation-reduction induction, hormonal status disruption and

different gene expression pathways of blastocyst-derived cells.

## Materials and Methods

### Chemicals

Bisphenol-S (99%, 4, 4'-Sulfonyldiphenol) (CAS No. 80-09-1) and ethanol (ACS grade; CAS No. 64-17-5) were purchased from Sigma-Aldrich Company (St. Louis, MO, USA). The stock solution was prepared by sufficient BPS powder dissolved in 100% ethanol, then, diluted in saline with less than 0.1-0.5% concentration of ethanol in the final solution (13).

### Animals

All adult, cycling, female NMRI mice (5-6 weeks), which were in the estrous stage, were purchased and maintained in Urmia university animal house and the estrous cycle animals were chosen by vaginal smear (39). After seven days of adaptation, they were housed in standard glass cages with a 12-hr-12 hr light /dark cycle. Animals were fed phytoestrogens free diet with free access to glass water bottles.

### Experimental design

All steps were performed based on the National Institutes of Health Guide for Care and Use of Laboratory Animals and followed the principles of "Use of Animals in Toxicology", with a slight modification, which were approved by Animal Ethics Committee (AEC) of University of Urmia (37). Briefly (Figure 1), different doses of BPS (0, 1, 5, 10, 50 and 100  $\mu\text{g}/\text{kg}$  bodyweight/ day) were administered subcutaneously (SC) for eliminating the first-pass effect and increasing serum unconjugated to conjugated BPS ratio more than oral absorption (10) for 21 consecutive days. Three dosages lower than 50  $\mu\text{g}/\text{kg}$  were selected according to the safe dose of BPA/day that was announced by the US Environmental Protection Agency (EPA) (40).

One day after the last administration, blood samples were directly collected from hearts of the five mice of each group and centrifuged at 12000  $\times g$  for 10 min; then, mice were euthanized using Ketamine/Xylazine (45 mg/kg; 35 mg/kg; IP) (41) and ovarian tissue samples were harvested, which were immediately placed in

**Table 1.** Details of primers (Primer sequence, amplicon size, and accession number) of different genes used in RT-PCR

Primer	Primer sequence (5'-3')	Amplicon size	Reference and/or gene accession number
P53 (F)	5' CCCGAGTATCTGGAAGACAG 3'	769 bp	NM_011640.3
P53 (R)	5' ATAGGTCGGCGGTTTCAT 3'		
Bax (F)	5' CGGCGAATTGGAGATGAACTG 3'	169 bp	NM_007527.3
Bax (R)	5' GCAAAGTAGAAGAGGGCAACC 3'		
ER $\alpha$ (F)	5' CACGATTGATAAAAACAGGAGGAA 3'	108 bp	NM_001001443
ER $\alpha$ (R)	5' CTCCTCCTCTTCGGTCTTTC 3'		
ER $\beta$ (F)	5' CATTGCCAGCCGTCAGTTCT 3'	120 bp	NM_174051
ER $\beta$ (R)	5' GCTCCCACTAGCCTTCCTTTTC 3'		
Nrf2 (F)	5' CGGTATGCAACAGGACATTG 3'	263 bp	NM_172086
Nrf2 (R)	5' ACTGGTTGGGGTCTTCTGTG 3'		
E2f1 (F)	5' TTGACCCCTCTGGATTCTCG 3'	198 bp	NM_007891.5
E2f1 (R)	5' CCCTTTGGTCTGCTCAATGT 3'		
18 SrRNA (F)	5' GCAATTATCCCATGAACG 3'	123 bp	NR_003278
18 SrRNA (R)	5' GGCCTCACTAAACATCCAA 3'		

phosphate buffer saline. The remaining mice were super-ovulated (5 mice per group) by pregnant mare's serum gonadotropin (PMSG) and human chorionic gonadotropin (hCG) as described before (37). All oocytes were collected and inseminated by one normal mouse capacitated sperm. Four hours after insemination, all fertilized zygotes was transferred to the HTF medium and Blastocyst (10-20 in each group) was collected four days later. All samples were preserved in a refrigerator at -20 °C for further evaluation.

#### Hormone level analysis

Follicle-stimulating hormone (FSH) and luteinizing hormone (LH) level in serum were measured by Enzyme-Linked Immunosorbent Assay (ELISA) as described in the instructions provided by the kit manufacturer (Pishtaz Teb Diagnostics., Iran).

#### Total RNA isolation and cDNA preparation

Total RNA extraction was performed by the Choi method with slight modifications (42). Three pools of 20 unhatched (early blastocyst, blastocyst and expanded blastocyst stage) embryos were used, which were collected on the fourth day following *in vitro* fertilization (IVF). Subsequently, all embryos were lysed by RNX plus solution (Cinnagen, Iran) according to the manufacturer's procedure. 1 ml of RNX solution was used to homogenize embryos and it was left intact at room temperature for 5 min. Later, 200  $\mu$ l chloroform was added to the tube and centrifuged for 15 min at the rate of 12000 g at 4 °C. RNA was obtained from the solution and an equal amount of isopropanol was added to it. The resulting compound was centrifuged again by the same procedure for 10 min. 75% ethanol was used to rinse the pellet and it was suspended again in 50  $\mu$ l of Diethylpyrocarbonate (DEPC) treated water. Finally, total RNA concentration was examined by spectrophotometry (Thermo Scientific NanoDrop ND-2000 UV spectrophotometer,) at OD  $\leq$  1.9, 260/280 ratio. The RNA quantity and integrity were examined by the absorbance ratio A260/A280 nm and 1% agarose gel electrophoresis, respectively.

Equal RNA concentration (1  $\mu$ g of DNase-treated total RNA) of each group was used for cDNA synthesis. cDNA Synthesis Kit Revert Aid was obtained from the

Fermentas Corporation (Germany). One  $\mu$ g of RNA was reverse transcribed with 5X Reaction Buffer, 20 U/ $\mu$ l Ribolock RNase inhibitor, 10 mM dNTP, 200 U/ $\mu$ l MMLV reverse transcriptase, and oligo (dt)18 primer in a 20  $\mu$ l reaction in order to synthesize cDNA. The resulting combination was incubated for 60 min at 42 °C. Later, the enzyme was inactivated for 5 min at 70 °C.

#### Real-time PCR (RT-PCR) of different gene pathways of blastocyst-derived cells

The integrated cDNA of all groups was further used in RT-PCR to reveal changes in the expression of the estrogen signaling pathway (ER $\alpha$  and ER $\beta$ ) and apoptosis induction pathway (P53, Bax, Nrf2 and E2F1) genes. The details of gene specific sequence primers are shown in Table 1. All samples were run in triplicate and 18SrRBA was chosen as a house-keeping gene to normalize the input load of cDNA between samples. Hot Taq Eva Green (cat# BT 11101- SinaClon, Iran) chemistry was used for sample preparation. The reaction's final volume equaled 25  $\mu$ l, which included 1  $\mu$ l of cDNA mixed with 1  $\mu$ M primers (forward and reverse) for each gene. Thermal cyclings used for the reaction were as follows: initial denaturation at 95 °C for 15 min, 40 cycles of denaturation at 95 °C for 15 sec, primer annealing and extension at 62 °C for 1 min. The fluorescence was seized at the end of the extension step. At the end of each round, a melt-curve analysis was performed to conclude the specificity of the amplification. The relative expression of each gene was examined using the delta CT model. For product homogeneity certainty, the melting curve analysis was assessed after the real-time PCR procedure.

#### Examination of ovarian tissue oxidative activity

Both harvested ovarian tissues were homogenized using a soft tissue homogenizer (Omni International, USA) in 9 volumes of ice-cooled 0.9% buffer saline solution. The homogenate was then centrifuged at 4000 x g for 10 min at 4 °C and the supernatant was used for estimating antioxidant capacity. Ferric reduction antioxidant power assay was used for Total Antioxidant Capacity (TAC) assay according to Benzie and Strain's study (43). Moreover, other anti-oxidant enzymes' activities including superoxide dismutase (SOD), glutathione peroxidase (GPx), and CAT were determined

**Table 2.** Bisphenol-S (BPS) dose dependent effect on hormone status

Groups	FSH mIU/MI	LH mIU/MI
control	3.66±1.25 <sup>a</sup>	2.71±1.07 <sup>a</sup>
B0	3.58±1.33 <sup>a</sup>	2.79±1.45 <sup>a</sup>
B1	3.30±0.91 <sup>b</sup>	2.25±0.64 <sup>b</sup>
B5	2.76±1.14 <sup>c</sup>	1.88±1.09 <sup>c</sup>
B10	1.97±0.76 <sup>d</sup>	1.10±0.77 <sup>d</sup>
B50	1.47±0.61 <sup>e</sup>	1.20±0.67 <sup>e</sup>
B100	1.20±0.89 <sup>f</sup>	0.83±0.73 <sup>e</sup>

Different letters indicate significant differences between groups. Values represent means±SEM

BPS: Bisphenol-S; FSH: Follicle-stimulating hormone; LH: luteinizing hormone

as described by Nishikimi *et al.* Paglia and Valentine, and Sinha, respectively (44-46) and the activity of MDA was determined in our previous study (37).

### Statistical analysis

All analyses were performed using SPSS ver.24 (SPSS, Chicago, IL, USA) and data was stated as the mean±Standard Error (SE). Welch- one-way analysis of variance (ANOVA) test was used for comparison between groups and possible correlation was assessed by the Spearman test.  $P<0.05$  was considered significant in all tests. Also multiple linear regression and stepwise regression were performed in order to determine the main indicator. Hatch percentage was selected as the dependent variable and all other variables including oxidative stress, hormone status and gene expression were chosen as predictors.

## Results

Note: There were not any malformation or organelle differences during the evaluation, hence, no animal was

eliminated during the experiment.

To investigate the dose-dependent influence of BPS on hormone levels, LH and FSH status were examined in which LH and FSH reduction was induced at the lowest administered doses. As represented in Table 2, 1 µg/kg of BPS caused a significant reduction in hormone status; interestingly, 50 µg/kg and 100 µg/kg had the same effect on LH and FSH hormone levels. Moreover, there were meaningful differences in LH and FSH status between two high dosage groups.

As shown in Table 3, BPS induced oxidative stress in ovarian tissue even in the lowest dosage. All factors, except GPx, indicated significant oxidative stress in ovarian tissue induced by exposure to lowest BPS doses. Comparing the highest doses of BPS (50 µg/kg and 100 µg/kg), there were meaningful differences between all groups and only TAC status was constant.

Total RNA concentration detected by spectrophotometer ranged from 80 to 132 ng/µl for gene expression analysis. The Integrity and quality of all primers were examined by PCR which was verified by appearance of 769 bp, and 160 bp and absence of 1.3 kb DNA band on 2% agarose gel (Figure 2). In the present study, a different concentration of BPS was set for analysis of *P53*, *Bax*, estrogen receptor alpha (*ERα*), *ERβ*, *E2f1* and *Erf2* genes by RT-PCR. It was found that expressions of *P53*, *Bax*, *ERα*, *ERβ* and *E2f1* genes were significantly higher ( $P<0.05$ ) in doses of more than 5 µg/kg as compared to control groups. In comparison between 100 µg/kg and 50 µg/kg groups, there were significant differences in gene representation except for *Nrf2*. Interestingly, down-regulation was observed in *P53* and *F2f1* status in 100 µg/kg and 50 µg/kg groups which might indicate BPS-induced apoptosis pathways' exchange from *P53* dependent to *p53* independent pathway in doses more than 50 µg/kg (Table 4).

**Table 3.** Different bisphenol-S (BPS) dose dependent oxidative induction in ovarian tissue

Groups	CAT (U/mg protein)	GPx (U/mg protein)	SOD (U/mg protein)	TAC (µMol/l)
Control	21.39±1.27 <sup>a</sup>	1.23±0.34 <sup>a</sup>	2.61±0.36 <sup>a</sup>	2.38±1.32 <sup>a</sup>
B0	20.07±1.44 <sup>ab</sup>	1.22±0.30 <sup>a</sup>	2.58±0.28 <sup>a</sup>	2.35±0.89 <sup>a</sup>
B1	18.53±1.09 <sup>b</sup>	1.18±0.18 <sup>a</sup>	2.18±0.30 <sup>b</sup>	1.96±1.07 <sup>b</sup>
B5	15.28±1.35 <sup>c</sup>	1.03±0.24 <sup>b</sup>	2.12±0.18 <sup>bc</sup>	1.63±0.93 <sup>c</sup>
B10	11.74±0.87 <sup>d</sup>	0.86±0.27 <sup>c</sup>	1.98±0.27 <sup>c</sup>	1.09±0.81 <sup>d</sup>
B50	9.6±1.79 <sup>e</sup>	0.79±0.18 <sup>c</sup>	1.75±0.33 <sup>d</sup>	1.17±0.49 <sup>e</sup>
B100	6.07±1.04 <sup>f</sup>	0.61±0.21 <sup>d</sup>	1.48±0.29 <sup>e</sup>	1.27±0.50 <sup>e</sup>

SOD: Superoxide dismutase; GPx: Glutathione peroxidase; CAT: Catalase; TAC: Total Anti-oxidant Capacity; BPS: Bisphenol-S; superoxide dismutase

Different letters indicate significant differences between groups. Values represent means±SEM

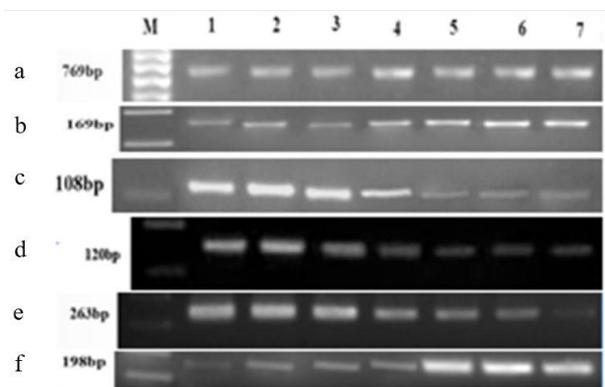
**Table 4.** Bisphenol-S (BPS) different dosage effect on level of mRNA expressions of genes (fold-change)

Groups	P53	Bax	ERα	ERβ	Nrf2	E2f1
Control	0.51±0.34 <sup>a</sup>	0.43±0.15 <sup>a</sup>	0.94±0.24 <sup>a</sup>	0.36±0.16 <sup>a</sup>	0.43±0.11 <sup>a</sup>	0.37±0.19 <sup>a</sup>
B0	0.50±0.22 <sup>a</sup>	0.46±0.24 <sup>a</sup>	0.93±0.19 <sup>a</sup>	0.35±0.20 <sup>a</sup>	0.42±0.20 <sup>a</sup>	0.39±0.08 <sup>a</sup>
B1	0.55±0.16 <sup>a</sup>	0.51±0.27 <sup>ab</sup>	0.89±0.22 <sup>a</sup>	0.42±0.17 <sup>ab</sup>	0.46±0.24 <sup>a</sup>	0.44±0.24 <sup>a</sup>
B5	0.83±0.25 <sup>b</sup>	0.58±0.19 <sup>b</sup>	0.78±0.13 <sup>b</sup>	0.47±0.11 <sup>b</sup>	0.52±0.15 <sup>a</sup>	0.57±0.17 <sup>b</sup>
B10	0.89±0.19 <sup>bc</sup>	0.89±0.20 <sup>c</sup>	0.57±0.18 <sup>c</sup>	0.53±0.24 <sup>bd</sup>	1.08±0.23 <sup>b</sup>	0.73±0.25 <sup>c</sup>
B50	1.04±0.33 <sup>c</sup>	0.96±0.29 <sup>c</sup>	0.30±0.25 <sup>d</sup>	0.58±0.27 <sup>c</sup>	1.57±0.09 <sup>c</sup>	1.09±0.13 <sup>d</sup>
B100	1.027±0.18 <sup>d</sup>	1.41±0.22 <sup>d</sup>	0.26±0.14 <sup>d</sup>	0.69±0.18 <sup>d</sup>	1.60±0.31 <sup>c</sup>	0.92±0.26 <sup>c</sup>

Different letters indicate significant differences between groups. Values represent means±SEM

**Table 5.** Spearman's correlation coefficient between the results of the all experimental on blood, ovary tissue and oocytes of mice

	MDA	TAC	CAT	GPx	SOD	LH	FSH	P53	Bax	ERα	ERβ	Nrf2	E2F1	Fertilization rate	Two cell	Blastocyst	Hatch
MDA	1																
TAC		1															
CAT			1														
GPx				1													
SOD					1												
LH						1											
FSH							1										
P53								1									
Bax									1								
ERα										1							
ERβ											1						
Nrf2												1					
E2F1													1				
Fertilization Rate														1			
Two. Cell															1		
Blastocyst																1	
Hatch																	1



**Figure 2.** Gene's bands derived from cDNAs samples PCR using a) P53 (769 bp), b) Bax (169 bp), c) ERα (108 bp), d) ERβ (120 bp), e) Nrf2 (263 bp), and f) E2f1 (198 bp) gene primer; Lane M: 100 bp DNA ladder; Lane 1, control; Lane 2, B0; Lane 3, B1; Lane 4, B5; Lane 5, B10; Lane 6, B50; Lane 7, B100 groups

As seen in Table 5, all data had significant correlation with each other. A significant positive correlation was detected between all IVF parameters and all anti-oxidative enzymes except MDA. Significant negative correlation was observed between all examined genes and IVF parameters except ERα. A positive correlation was detected between LH, FSH, ERα, and all IVF parameters.

In multiple regression using ovarian oxidative stress, hormonal status and gene expression as predictors

of IVF outcome, Nrf2, E2f1, MDA, and CAT were significantly associated with Hatch percentage (Table 6). The latter parameters' predictive values were established by multiple logistic regression analysis and it was revealed that Nrf2, E2f1, MDA, and CAT could be used as indicators for IVF outcome prediction (Table 6).

### Discussion

Increasing use of BPS in new products raised public anxiety about its detrimental impact on human health. Recently, BPS was detected in blood and urinary samples of human beings with various ages (19, 47-49). Recent research indicates that after BPA, BPS is chosen as the second compound of bisphenol family which has the highest dermal absorption level from thermal paper (50). This study was conducted to illustrate the effects of exposure to the low concentrations of BPS on hormone disruption and oxidative factor status in mature females. It also aimed to define the exact mechanisms involved in next generation blastocyst-derived cells by evaluation of different gene expression pathways.

The levels of reproductive hormones (LH, FSH, testosterone, progesterone, and estrogen) have been reported to be changed by BPA exposure in different doses, length and method of administration (35, 51-55). Furthermore, the toxic effect of BPA on postnatal development was evaluated with an insignificant change of sex hormones (56). Based on our experiment, even at its lowest doses, BPS impaired LH and FSH production and there were significant differences between high-

**Table 6.** Multiple linear and stepwise linear regression between ovarian oxidative stress, hormonal status and gene expression with Hatch percentage

Dependent Variable	Regression					Stepwise Regression					
	Unstandardized	Coefficients	Standardized coefficients	t	P. value	Unstandardized	Coefficients	Standardized coefficients	t	P. value	
Hatch	Nrf2	-11.967	1.807	-7.60	-6.623	.000	-11.677	1.601	-.741	-7.295	.000
	E2f1	-11.316	2.265	-.395	-4.996	.000	-11.226	2.101	-.391	-5.342	.000
	CAT	-.511	.199	-.343	-2.562	.012	-.522	.139	-.351	-3.757	.000
	MDA	-.720	.288	-.202	-2.497	.014	-.602	.268	-.169	-2.248	.027
	GPx	-2.801	2.124	-.092	-1.319	.191					
	FSH	.624	.384	.095	1.624	.108					
	P53	2.075	1.603	.072	1.295	.199					
	Bax	1.437	1.770	.065	.812	.419					
	Era	.338	2.101	.013	.161	.873					
	ERb	1.570	2.721	.031	.577	.565					

dose groups (50 µg/kg and 100 µg/kg), and it should be noted that 100 µg/kg of BPS induced higher reduction compared with other groups. Studies revealed that administering higher doses of BPS led to reduction of LH and FSH in female mice and this result is in agreement with the results obtained in this study (57). However, further studies are required for investigating the exact effect of BPS on hormonal production mechanisms. It should be noted that the effect of BPA on LH/FSH was previously established (35, 54, 55). Based on previous studies and the current study, it is suggested that exposure to low doses of BPS might have disrupted sex hormone secretion, production and performance potential in different animals and organelles.

Interestingly, the effects of BPA and other members of the Bisphenol family on ROS production in different tissues were previously evaluated (58-60). The findings of our study also indicated that BPS increased oxidative stress status in ovarian tissue even in low doses by determining different antioxidant enzyme parameters including SOD, GSH-Px, CAT as enzymatic antioxidant and TAC as a non-enzymatic antioxidant detector. Besides, our previous study indicated that BPS increases lipid peroxidation, which is in line with the results of the present study (37). In addition, recent studies indicated that BPS increases oxidation potential in testis tissue and short-term BPS cultured spermatozoa (13, 33). Furthermore, oxidation induction of BPA and BPS might be related to their estrogenic potential especially because of their strong effect on estrogen receptor alpha (25, 61-63). The correlation between hormonal status and ovarian tissue oxidative stress delineated that BPS not only influences ovarian tissue as an oxidant, but also mimics the estrogenic effect that interrupts hormonal activity in the body.

The effect of BPA on pro-apoptotic and apoptosis gene expressions such as Bax and P53 as main indicators were previously established. Bax expression was up-regulated in response to BPA exposure in granulosa cells (64) and P53 expression was induced following 24hr BPA subjection in the ovarian cancer cell line (65). Furthermore, a correlation between BPA estrogen receptor stimulation and apoptosis induction was demonstrated by assessment of apoptotic genes in different *in vitro* cellular cultures (35, 36, 66). In this study, dose-dependent manner of p53 and Bax gene increase was induced in blastocyst collected from females exposed to different doses of BPS (Table 4). In agreement with our study, an *in vitro* study demonstrated that 40 mM of BPS in human adrenal cortico-carcinoma cells promotes P53 growth and apoptosis (67). Additionally, significant BPS dose-dependent apoptosis and DNA damage induction were detected in human testis tissue and human hepatoma cell line, respectively (61, 68), which confirms the results of the present study. P53 exact mechanism through which it influences embryos was previously explained. Briefly, the increase in P53 gene expression during early embryonic development induced Bax expression, as a result, it promotes apoptosis in the early stages of development. Consequently, embryonic arrest after apoptosis initiation could lead to embryonic lethality and abnormal embryo expansion (69). Additionally, embryonic hatch rate reduction and IVF outcome

decrease were observed as mentioned before (37).

Estrogen Receptors (ERs) are involved in the early stages of fetal development and embryonic organ differentiation (70, 71). The increase in ERs has been demonstrated in brains exposed to BPA (72), and the results indicate that BPA influences ERα gene expression and promotes apoptosis in breast cancer cells (66). A previous study demonstrated that in utero exposure of BPA during pregnancy influences ERα more than ERβ by mimicking estrogenic effect, therefore, it leads to adverse effects observed in early stages of development (73). Also, steroidogenic enzyme activity was increased by BPS prenatal exposure in next male generation testis and 50 µg/kg dosage group showed significant alteration compared to lower dosage in estradiol status and steroidal gene expression (34). Our study demonstrated that the effect of BPS on blastocyst-derived cells revealed P53 and ER expression and decreased ERα/ERβ ratio in response to the dose-dependent manner of BPS. Although, ER does not express after zygote fertilization until the blastocyst stage, it has a crucial role in blastocyst differentiation and implantation delay rate (74, 75). In agreement with our results, acute administration of BPS involved ER pathways especially primary macrophages and embryo-larvae in fish, however, up-regulation was detected in both ER genes expression (76, 77). However, dose and long-term response manner of BPS showed down-regulation in ERα expression in cell culture (78). Indeed, ERβ compared to ERα involved in long-term effects of ERs, and it elucidates the main cause of ERα reduction in the present study (77). It also suggested BPS selective specificity to ERα which significantly influences receptors in high dosage (79). Moreover, consistent with our study, long-term BPS administration down-regulated ERα expression in the ovary (21).

The E2f transcription factor family, is known as an essential part in the regulation of cell proliferation by mitosis checkpoint between G1/S phase (65), apoptosis induction in P53 dependent and independent pathway, and development in pre-implantation embryos (80, 81). E2f1 has been suggested to participate as the main p53-related apoptosis inducer (82-85), developmental stages of early embryos (80), and DNA replication malfunction in ovarian follicular and embryonic stem cells (86, 87). Based on the aforementioned studies, it is suggested that E2f1 has a crucial role in ovarian cancer and apoptosis. Our study showed that E2F1 gene expression was promoted in blastocyst cells in different BPS groups, which might indicate BPS teratogenicity and apoptosis-induced potential for infants even in low doses of administration. Interestingly, down-regulation was observed in the comparison of 50 and 100 groups and it probably indicates apoptosis pathway exchange from P53 dependent to P53 independent. Moreover, the same down-regulation was detected in P53 gene expression that is confirmed by pathway exchange hypothesis in 50 µg/kg. Our results are consistent with the study conducted by Cheraghi *et al.* who examined Zearalenone's influence on E2F1 expression in which gene expression status was detected in moderate dosage (2 mg/kg) to be more than high dosage (4 mg/kg) in testis. They suggested reduction of E2f1 in the middle dose might be due to the apoptosis pathway exchange from p53 dependent to independent pathways (84).

Possible side effects of low dosages of BPA (10 nM) were determined in ER $\alpha$  negative breast cell culture which apoptosis induced by E2f1 dependent pathway. Also, a time-dependent manner and different cell culture examination showed the same results and suggested a possible ER independent side effect of BPA (88).

Interestingly, the correlation between ER and E2F1 in breast cancer cell line is a controversial issue. Some studies demonstrated a repressive effect between ER stimulation, especially ER $\alpha$ , and E2F1 gene expression (89), in contrast, stimulation of E2f1 expression was observed by ER $\alpha$  in normal MCF-7 breast cancer cell lines (90, 91) and cell lines that were tolerated to Tamoxifen (92). The correlation between ER and E2f1 was detected in our study, and positive correlation was detected between ER $\beta$  and E2F1, but negative correlation was detected between ER $\alpha$  and E2F1. Due to positive correlation between E2F1 and P53 as well as previous investigations, it could be suggested that BPS administration induced apoptosis in blastocyst by N2f1/p53 dependent pathway which is in agreement with Caspase 8 detection in cell culture exposed to different BPS concentrations (10<sup>-14</sup> M and 10<sup>-8</sup> M) (78). To the best of our knowledge, this is the first study to demonstrate the involvement of ER and E2f1 in different apoptosis induction pathways in blastocyst-derived cells that are harvested from BPS-exposed mother.

Nuclear factor erythroid 2-related factor 2 (Nrf2) is another transcription factor involved in embryonic development and mitosis entry. Nrf2 has multifunctional subgroup genes, that participate in cytoplasmic defense against many environmental oxidants (antioxidant proteins) such as SOD, CAT, cell injuries by cyclin B-CDK1 complex and phase 2 detoxifying enzymes (93-96). Many previous studies presumed Nrf2 as an oxidant status indicator in pre-implantation embryos especially in blastocyst stage (93, 97, 98). BPA's effect on lipid accumulation in the liver during pregnancy and in offspring was examined by Nrf2 as a pro-lipogenic factor. As a result, an increase in Nrf2 was reported both in pregnant mice exposed to BPA and their next generation offspring (99). The results of this study demonstrated that BPS induced significant oxidative stress that was indicated by Erf2 up-regulation in dosage higher than 10  $\mu$ g/kg. Furthermore, no significant difference was observed between 50 and 100  $\mu$ g/kg groups of BPS. A negative strong correlation was detected between Nrf2 and GPx, SOD, and Catalase which is consistent with the aforementioned hypothesis that signified Nrf2 as an indicator of oxidative stress.

Stepwise regression analysis revealed that E2f1, Nrf2, CAT, and MDA could be chosen as predictors of hatch percentage (Table 6) and current regression analysis failed to determine the meaningful relationship between other parameters with hatch capability. However, the Spearman coefficient determined meaningful correlation between all parameters and hatch percentages.

## Conclusion

The current study illustrates that low concentration of BPS had a detrimental effect on the female reproductive system by oxidative stress induction, gonadal hormone reduction, and next-generation apoptosis induction,

which was determined by examination of Blastocyst-derived cells. Also, E2f1, Nrf2, CAT, and MDA have predictive value in the hatch blastocyst stage which can be used for pre-implantation embryo quality examination. Further studies are required to determine the exact mechanisms of BPS hormonal status as well as determining BPS capability for induced teratogenic effect on the next generations.

## Acknowledgment

The authors would like to sincerely thank the Faculty of Veterinary Medicine and Urmia University Research Council for the approval and financial support of this research. The results presented in this paper were part of a student thesis.

## Conflicts of Interest

The authors declare that there are no conflicts of interest.

## References

- Sengupta P, Banerjee R. Environmental toxins: Alarming impacts of pesticides on male fertility. *Hum Exp Toxicol* 2014; 33:1017-1039.
- Maqbool F, Mostafalou S, Bahadar H, Abdollahi M. Review of endocrine disorders associated with environmental toxicants and possible involved mechanisms. *Life Sci* 2016; 145:265-273.
- Hiroi H, Tsutsumi O, Momoeda M, Takai Y, Osuga Y, Taketani Y. Differential interactions of bisphenol A and 17 $\beta$ -estradiol with estrogen receptor  $\alpha$  (ER $\alpha$ ) and ER $\beta$ . *Endocrine J* 1999; 46:773-778.
- Guignard D, Gayraud V, Lacroix M, Puel S, Picard-Hagen N, Viguié C. Evidence for bisphenol A-induced disruption of maternal thyroid homeostasis in the pregnant ewe at low level representative of human exposure. *Chemosphere* 2017; 182:458-467.
- Park C, Choi W, Hwang M, Lee Y, Kim S, Yu S, et al. Associations between urinary phthalate metabolites and bisphenol A levels, and serum thyroid hormones among the Korean adult population-Korean National Environmental Health Survey (KoNEHS) 2012-2014. *Sci Total Environ* 2017; 584:950-957.
- Leranth C, Hajszan T, Szigeti-Buck K, Bober J, MacLusky NJ. Bisphenol A prevents the synaptogenic response to estradiol in hippocampus and prefrontal cortex of ovariectomized nonhuman primates. *Proc Natl Acad Sci* 2008; 105:14187-14191.
- Liang H, Xu W, Chen J, Shi H, Zhu J, Liu X, et al. The association between exposure to environmental Bisphenol-A and gonadotropic hormone levels among men. *PloS one* 2017; 12: e0169217.
- Ehrlich S, Williams PL, Missmer SA, Flaws JA, Berry KF, Calafat AM, et al. Urinary bisphenol A concentrations and implantation failure among women undergoing *in vitro* fertilization. *Environmental health perspectives. Environ Health Persp* 2012; 120:978-983.
- Ramakrishnan S, Wayne NL. Impact of bisphenol-A on early embryonic development and reproductive maturation. *Reprod Toxicol* 2008; 25:177-183.
- Alonso-Magdalena P, Garcia-Arévalo M, Quesada I, Nadal Á. Bisphenol-A treatment during pregnancy in mice: a new window of susceptibility for the development of diabetes in mothers later in life. *Endocrinology* 2015; 156:1659-1670.
- Cao Y, Qu X, Ming Z, Yao Y, Zhang Y. The correlation between exposure to BPA and the decrease of the ovarian reserve. *Int J*

- Clin Exp Pathol 2018; 11:3375-3382.
12. Zhang R, Liu R, Zong W. Bisphenol S Interacts with Catalase and Induces Oxidative Stress in Mouse Liver and Renal Cells. *J Agr Food Chem* 2016; 64:6630-6640.
  13. Ullah H, Jahan S, Ain QU, Shaheen G, Ahsan N. Effect of bisphenol S exposure on male reproductive system of rats: A histological and biochemical study. *Chemosphere* 2016; 152:383-391.
  14. Kabuto H, Amakawa M, Shishibori T. Exposure to bisphenol A during embryonic/fetal life and infancy increases oxidative injury and causes underdevelopment of the brain and testis in mice. *Life Sci* 2004; 74:2931-2940.
  15. Kadasala NR, Narayanan B, Liu Y. International trade regulations on BPA: Global health and economic implications. *Asian Dev Pol Rev* 2016; 4:134-142.
  16. Del Moral LI, Le Corre L, Poirier H, Niot I, Truntzer T, Merlin J-F, et al. Obesogen effects after perinatal exposure of 4, 4'-sulfonyldiphenol (Bisphenol S) in C57BL/6 mice. *Toxicology* 2016; 357-358:11-20.
  17. Wan Y, Xia W, Yang S, Pan X, He Z, Kannan K. Spatial distribution of bisphenol S in surface water and human serum from Yangtze River watershed, China: Implications for exposure through drinking water. *Chemosphere* 2018; 199:595-602.
  18. Thayer KA, Taylor KW, Garantziotis S, Schurman SH, Kissling GE, Hunt D, et al. Bisphenol A, bisphenol S, and 4-hydroxyphenyl 4-isopropoxyphenylsulfone (BPSIP) in urine and blood of cashiers. *Environ Health Perspect* 2016; 124:437-444.
  19. Liao C, Liu F, Alomirah H, Loi VD, Mohd MA, Moon HB, et al. Bisphenol S in urine from the United States and seven Asian countries: occurrence and human exposures. *Environ Sci Technol* 2012; 46:6860-6866.
  20. Rezg R, Abot A, Mornagui B, Aydi S, Knauf C. Effects of Bisphenol S on hypothalamic neuropeptides regulating feeding behavior and apelin/APJ system in mice. *Ecotox Environ Safe* 2018; 161:459-466.
  21. Hill CE, Sapouckey SA, Suvorov A, Vandenberg LN. Developmental exposures to bisphenol S, a BPA replacement, alter estrogen-responsiveness of the female reproductive tract: A pilot study. *Cogent Med* 2017; 4:1317690.
  22. Catanese MC, Vandenberg LN. Bisphenol S (BPS) alters maternal behavior and brain in mice exposed during pregnancy/lactation and their daughters. *Endocrinology* 2016; 158:516-530.
  23. LaPlante CD, Catanese MC, Bansal R, Vandenberg LN. Bisphenol S alters the lactating mammary gland and nursing behaviors in mice exposed during pregnancy and lactation. *Endocrinology* 2017; 158:3448-3461.
  24. Kinch CD, Ibhazehiebo K, Jeong J-H, Habibi HR, Kurrasch DM. Low-dose exposure to bisphenol A and replacement bisphenol S induces precocious hypothalamic neurogenesis in embryonic zebrafish. *Proc Natl Acad Sci* 2015; 112:1475-1480.
  25. Qiu W, Zhao Y, Yang M, Farajzadeh M, Pan C, Wayne NL. Actions of bisphenol A and bisphenol S on the reproductive neuroendocrine system during early development in zebrafish. *Endocrinology* 2015; 157:636-647.
  26. Rochester JR, Bolden AL. Bisphenol S and F: a systematic review and comparison of the hormonal activity of bisphenol A substitutes. *Environ Health Perspect* 2015; 123:643-650.
  27. Ji K, Hong S, Kho Y, Choi K. Effects of bisphenol S exposure on endocrine functions and reproduction of zebrafish. *Environ Sci Technol* 2013; 47:8793-8800.
  28. Wang W, Zhang X, Wang Z, Qin J, Wang W, Tian H, et al. Bisphenol S induces obesogenic effects through deregulating lipid metabolism in zebrafish (*Danio rerio*) larvae. *Chemosphere* 2018; 199:286-296.
  29. Liu J, Li J, Wu Y, Zhao Y, Luo F, Li S, et al. Bisphenol A metabolites and bisphenol S in paired maternal and cord serum. *Environ Sci Technol* 2017; 51:2456-2463.
  30. Fol VL, Brion F, Hillenweck A, Perdu E, Bruel S, Ait-Aissa S, et al. Comparison of the *in vivo* biotransformation of two emerging estrogenic contaminants, BP2 and BPS, in zebrafish embryos and adults. *Int J Mol Sci* 2017; 18:704.
  31. Žalmanová T, Hošková K, Nevorál J, Adámková K, Kott T, Šulc M, et al. Bisphenol S negatively affects the meiotic maturation of pig oocytes. *Sci Rep* 2017; 7:1-1
  32. Ullah H, Ambreen A, Ahsan N, Jahan S. Bisphenol S induces oxidative stress and DNA damage in rat spermatozoa *in vitro* and disrupts daily sperm production *in vivo*. *Toxicol Environ Chem* 2017; 99:953-965.
  33. Shi M, Sekulovski N, MacLean JA, Hayashi K. Prenatal exposure to bisphenol A analogues on male reproductive functions in mice. *Toxicol Sci* 2018; 163:620-631.
  34. Ok S, Kang JS, Kim KM. Cultivated wild ginseng extracts upregulate the anti-apoptosis systems in cells and mice induced by bisphenol A. *Mol Cell Toxicol* 2017; 13:73-82.
  35. Yin L, Dai Y, Cui Z, Jiang X, Liu W, Han F, et al. The regulation of cellular apoptosis by the ROS-triggered *PERK/EIF2 $\alpha$ /chop* pathway plays a vital role in bisphenol A-induced male reproductive toxicity. *Toxicol Appl Pharm* 2017; 314:98-108.
  36. Nourian A, Soleimanzadeh A, Jalali AS, Najafi G. Effects of bisphenol-S low concentrations on oxidative stress status and *in vitro* fertilization potential in mature female mice. *Vet Res Forum* 2017; 8:341-345.
  37. Nevorál J, Kolínko Y, Moravec J, Zalmanova T, Hoskova K, Prokesova S, et al. Long-term exposure to very low doses of bisphenol S affects female reproduction. *Reproduction* 2018; 156:47-57.
  38. Byers SL, Wiles MV, Dunn SL, Taft RA. Mouse estrous cycle identification tool and images. *PLoS one* 2012; 7:e35538.
  39. Rubin BS. Bisphenol A: an endocrine disruptor with widespread exposure and multiple effects. *J Steroid Biochem Mol Biol* 2011; 127:27-34.
  40. Babaei M, Najafi G, Jalali AS, Behfar M. Effects of unilateral iatrogenic vas deferens trauma on fertility: An experimental *in vitro* fertilization mice model study. *Bull Emerg Trauma* 2015; 3:122-127.
  41. Choi BI, Harvey AJ, Green MP. Bisphenol A affects early bovine embryo development and metabolism that is negated by an oestrogen receptor inhibitor. *Sci Rep* 2016; 6:29318-29337.
  42. Benzie IF, Strain JJ. The ferric reducing ability of plasma (FRAP) as a measure of "antioxidant power": the FRAP assay. *Anal Biochem* 1996; 239:70-76.
  43. Nishikimi M, Rao NA, Yagi K. The occurrence of superoxide anion in the reaction of reduced phenazine methosulfate and molecular oxygen. *Biochem Biophys Res Co* 1972; 46:849-854.
  44. Paglia DE, Valentine WN. Studies on the quantitative and qualitative characterization of erythrocyte glutathione peroxidase. *J Lab Clin Med* 1967; 70:158-169.
  45. Sinha AK. Colorimetric assay of catalase. *Anal Biochem* 1972; 47:389-394.
  46. Pell T, Eliot M, Chen A, Lanphear BP, Yolton K, Sathyanarayana S, et al. Parental concern about environmental chemical exposures and children's urinary concentrations of phthalates and phenols. *J Pediatr* 2017; 186:138-144.
  47. Zhou X, Kramer JP, Calafat AM, Ye X. Automated on-line column-switching high performance liquid chromatography isotope dilution tandem mass spectrometry method for the quantification of bisphenol A, bisphenol F, bisphenol S, and 11 other phenols in urine. *J Chromatogr B* 2014; 944:152-156.
  48. Björnsdotter MK, de Boer J, Ballesteros-Gómez A. Bisphenol A and replacements in thermal paper: A review. *Chemosphere*

- 2017; 182:691-706.
49. Fernández M, Bourguignon N, Lux-Lantos V, Libertun C. Neonatal exposure to bisphenol a and reproductive and endocrine alterations resembling the polycystic ovarian syndrome in adult rats. *Environ Health Perspect* 2010; 118:1217-1222.
50. Jin P, Wang X, Chang F, Bai Y, Li Y, Zhou R, et al. Low dose bisphenol A impairs spermatogenesis by suppressing reproductive hormone production and promoting germ cell apoptosis in adult rats. *J Biomed Res* 2013; 27:135-144.
51. Vahedi M, Saeedi A, Poorbaghi SL, Sepehrimanesh M, Fattahi M. Metabolic and endocrine effects of bisphenol A exposure in market seller women with polycystic ovary syndrome. *Environ Sci Pollut Res* 2016; 23:23546-23550.
52. Ma S, Shi W, Wang X, Song P, Zhong X. Bisphenol-A exposure during pregnancy alters the Mortality and Levels of Reproductive Hormones and Genes in Offspring Mice. *BioMed Res Int* 2017; 2017.
53. Patel S, Brehm E, Gao L, Rattan S, Ziv-Gal A, Flaws JA. Bisphenol-A exposure, ovarian follicle numbers, and female sex steroid hormone levels: Results from a CLARITY-BPA study. *Endocrinology* 2017; 158:1727-1738.
54. Ogo FM, Siervo GE, Gonçalves GD, Cecchini R, Guarnier FA, Anselmo-Franci JA, et al. Low doses of bisphenol A can impair postnatal testicular development directly, without affecting hormonal or oxidative stress levels. *Reprod Fert Develop* 2017; 29:2245-2254.
55. Ahsan N, Ullah H, Ullah W, Jahan S. Comparative effects of Bisphenol S and Bisphenol A on the development of female reproductive system in rats; a neonatal exposure study. *Chemosphere* 2018; 197:336-343.
56. Yazdani M, Andresen AMS, Gjøen T. Short-term effect of bisphenol-a on oxidative stress responses in Atlantic salmon kidney cell line: a transcriptional study. *Toxicol Mech Method* 2016; 26:295-300.
57. Huc L, Lemarié A, Guéraud F, Héliers-Toussaint C. Low concentrations of bisphenol A induce lipid accumulation mediated by the production of reactive oxygen species in the mitochondria of HepG2 cells. *Toxicol in vitro* 2012; 26:709-717.
58. Maćczak A, Cyrkler M, Bukowska B, Michałowicz J. Bisphenol A, bisphenol S, bisphenol F and bisphenol AF induce different oxidative stress and damage in human red blood cells (*in vitro* study). *Toxicol in vitro* 2017; 41:143-149.
59. Desdoits-Lethimonier C, Lesné L, Gaudriault P, Zalko D, Antignac J, Deceuninck Y, et al. Parallel assessment of the effects of bisphenol A and several of its analogs on the adult human testis. *Hum Reprod* 2017; 32:1465-1473.
60. Mobley JA, Brueggemeier RW. Estrogen receptor-mediated regulation of oxidative stress and DNA damage in breast cancer. *Carcinogenesis* 2004; 25:3-9.
61. Nadal A, Fuentes E, Ripoll C, Villar-Pazos S, Castellano-Muñoz M, Soriano S, et al. Extranuclear-initiated estrogenic actions of endocrine disrupting chemicals: Is there toxicology beyond paracelsus? *J Steroid Biochem Mol Biol* 2018; 176:16-22.
62. Bredhult C, Bäcklin B-M, Olovsson M. Effects of some endocrine disruptors on the proliferation and viability of human endometrial endothelial cells *in vitro*. *Reprod toxicol* 2007; 23:550-559.
63. Ptak A, Wróbel A, Gregoraszczyk EL. Effect of bisphenol-A on the expression of selected genes involved in cell cycle and apoptosis in the OVCAR-3 cell line. *Toxicol lett* 2011; 202:30-35.
64. Sengupta S, Obiorah I, Maximov P, Curpan R, Jordan V. Molecular mechanism of action of bisphenol and bisphenol A mediated by oestrogen receptor alpha in growth and apoptosis of breast cancer cells. *Br J Pharmacol* 2013; 169:167-178.
65. Rosenmai AK, Dybdahl M, Pedersen M, Alice van Vugt-Lussenburg BM, Wedebye EB, Taxvig C, et al. Are structural analogues to bisphenol a safe alternatives? *Toxicol Sci* 2014; 139:35-47.
66. Fic A, Žegura B, Sollner Dolenc M, Filipič M, Peterlin Mašič L. Mutagenicity and DNA damage of bisphenol A and its structural analogues in HepG2 cells. *Arh Hig Rada Toksikol* 2013; 64:189-200.
67. Choi J, Donehower L. p53 in embryonic development: maintaining a fine balance. *Cell Mol Life Sci* 1999; 55:38-47.
68. Bondesson M, Hao R, Lin C-Y, Williams C, Gustafsson J-Å. Estrogen receptor signaling during vertebrate development. *Biochim Biophys Acta* 2015; 1849:142-151.
69. Lemmen JG, Broekhof JL, Kuiper GG, Gustafsson J-Å, Van Der Saag PT, Van Der Burg B. Expression of estrogen receptor alpha and beta during mouse embryogenesis. *Mech Dev* 1999; 81:163-167.
70. Ramos JG, Varayoud J, Kass L, Rodríguez H, Costabel L, Muñoz-de-Toro Mn, et al. Bisphenol A induces both transient and permanent histofunctional alterations of the hypothalamic-pituitary-gonadal axis in prenatally exposed male rats. *Endocrinology* 2003; 144:3206-3215.
71. Susiarjo M, Hassold TJ, Freeman E, Hunt PA. Bisphenol A exposure in utero disrupts early oogenesis in the mouse. *PLoS genet* 2007; 3:e5.
72. Hou Q, Gorski J. Estrogen receptor and progesterone receptor genes are expressed differentially in mouse embryos during preimplantation development. *Proc Natl Acad Sci* 1993; 90:9460-9464.
73. Saito K, Furukawa E, Kobayashi M, Fukui E, Yoshizawa M, Matsumoto H. Degradation of estrogen receptor  $\alpha$  in activated blastocysts is associated with implantation in the delayed implantation mouse model. *Mol Hum Reprod* 2014; 20:384-391.
74. Qiu W, Yang M, Liu S, Lei P, Hu L, Chen B, et al. Toxic effects of bisphenol S showing immunomodulation in fish macrophages. *Environ Sci Technol* 2018; 52:831-838.
75. Viñas R, Watson CS. Mixtures of xenoestrogens disrupt estradiol-induced non-genomic signaling and downstream functions in pituitary cells. *Environ Health* 2013; 12:26-36.
76. Li Y, Perera L, Coons LA, Burns KA, Ramsey JT, Pelch KE, et al. Differential *in vitro* biological action, coregulator interactions, and molecular dynamic analysis of Bisphenol-A (BPA), BPAF, and BPS ligand-ER $\alpha$  complexes. *Environ Health* 2018; 126:017012.
77. Palena A, Mangiacasale R, Magnano AR, Barberi L, Giordano R, Spadafora C, et al. E2F transcription factors are differentially expressed in murine gametes and early embryos. *Mech Dev* 2000; 97:211-215.
78. Liao Y, Du W. Rb-independent E2F3 promotes cell proliferation and alters expression of genes involved in metabolism and inflammation. *FEBS Open Bio* 2017; 7:1611-1621.
79. Yamasaki L, Bronson R, Williams BO, Dyson NJ, Harlow E, Jacks T. Loss of E2F-1 reduces tumorigenesis and extends the lifespan of Rb1 (+/-) mice. *Nat Genet* 1998; 18:360-364.
80. Ziebold U, Lee EY, Bronson RT, Lees JA. E2F3 loss has opposing effects on different pRB-deficient tumors, resulting in suppression of pituitary tumors but metastasis of medullary thyroid carcinomas. *Mol Cell Biol* 2003; 23:6542-6552.
81. Cheraghi S, Razi M, Malekinejad H. Involvement of cyclin D1 and *E2f1* in zearalenone-induced DNA damage in testis of rats. *Toxicol* 2015; 106:108-116.
82. Zhan L, Zhang Y, Wang W, Song E, Fan Y, Wei B. E2F1: a promising regulator in ovarian carcinoma. *Tumor Biol* 2016; 37:2823-2831.

83. Cayirlioglu P, Ward WO, Key SCS, Duronio RJ. Transcriptional repressor functions of *Drosophila* *E2F1* and *E2F2* cooperate to inhibit genomic DNA synthesis in ovarian follicle cells. *Mol Cell Biol* 2003; 23:2123-2134.
84. Roos W, Christmann M, Fraser S, Kaina B. Mouse embryonic stem cells are hypersensitive to apoptosis triggered by the DNA damage O6-methylguanine due to high *E2F1* regulated mismatch repair. *Cell Death Differ* 2007; 14:1422-1432.
85. Pfeifer D, Chung YM, Hu MC. Effects of low-dose bisphenol A on DNA damage and proliferation of breast cells: the role of c-Myc. *Environ Health Perspect* 2015; 123:1271-1279.
86. Liao XH, Lu DL, Wang N, Liu LY, Wang Y, Li YQ, et al. Estrogen receptor  $\alpha$  mediates proliferation of breast cancer MCF-7 cells via a *p21/PCNA/E2F1*-dependent pathway. *FEBS J* 2014; 281:927-942.
87. Wang W, Dong L, Saville B, Safe S. Transcriptional activation of *E2F1* gene expression by  $17\beta$ -estradiol in MCF-7 cells is regulated by NF-Y-Sp1/estrogen receptor interactions. *Mol Endocrinol* 1999; 13:1373-1387.
88. Frieze S, Lupien M, Silver PA, Brown M. CARM1 regulates estrogen-stimulated breast cancer growth through up-regulation of *E2F1*. *Cancer Res* 2008; 68:301-306.
89. Louie MC, McClellan A, Siewit C, Kawabata L. Estrogen receptor regulates *E2F1* expression to mediate tamoxifen resistance. *Mol Cancer Res* 2010; 8:343-352.
90. Lin Y, Sui LC, Wu RH, Fu HY, Xu JJ, Qiu XH, et al. Nrf2 inhibition affects cell cycle progression during early mouse embryo development. *J Reprod Develop* 2018;64:49-55.
91. Kobayashi M, Yamamoto M. Molecular mechanisms activating the *Nrf2-Keap1* pathway of antioxidant gene regulation. *Antioxid Redox Sign* 2005; 7:385-394.
92. Gad A, Hoelker M, Besenfelder U, Havlicek V, Cinar U, Rings F, et al. Molecular mechanisms and pathways involved in bovine embryonic genome activation and their regulation by alternative *in vivo* and *in vitro* culture conditions. *Biol Reprod* 2012; 87:100, 1-13.
93. Held E, Salilew-Wondim D, Linke M, Zechner U, Rings F, Tesfaye D, et al. Transcriptome fingerprint of bovine 2-cell stage blastomeres is directly correlated with the individual developmental competence of the corresponding sister blastomere. *Biol Reprod* 2012; 87:154, 1-13.
94. Bomfim MM, Andrade GM, del Collado M, Sangalli JR, Fontes PK, Nogueira MF, et al. Antioxidant responses and deregulation of epigenetic writers and erasers link oxidative stress and DNA methylation in bovine blastocysts. *Mol Reprod Dev* 2017; 84:1296-1305.
95. Amin A, Gad A, Salilew-Wondim D, Prastowo S, Held E, Hoelker M, et al. Bovine embryo survival under oxidative-stress conditions is associated with activity of the *NRF2*-mediated oxidative-stress-response pathway. *Mol Reprod Dev* 2014; 81:497-513.
96. Shimpi PC, More VR, Paranjpe M, Donepudi AC, Goodrich JM, Dolinoy DC, et al. Hepatic lipid accumulation and *Nrf2* expression following perinatal and peripubertal exposure to Bisphenol-A in a mouse model of nonalcoholic liver disease. *Environ Health Perspect* 2017; 125:087005.